

MasterClass 20



"Uso de fango aeróbico granular para el tratamiento de las aguas residuales industriales agroalimentarias"

Patrocinada por:



Aurora Rosa

Doctoranda FPU en el programa de doctorado de Ingeniería Civil de la Universidad de Granada en el Departamento de Microbiología

Grado en Biología y Máster en Investigación y Avances en Microbiología en la UGR





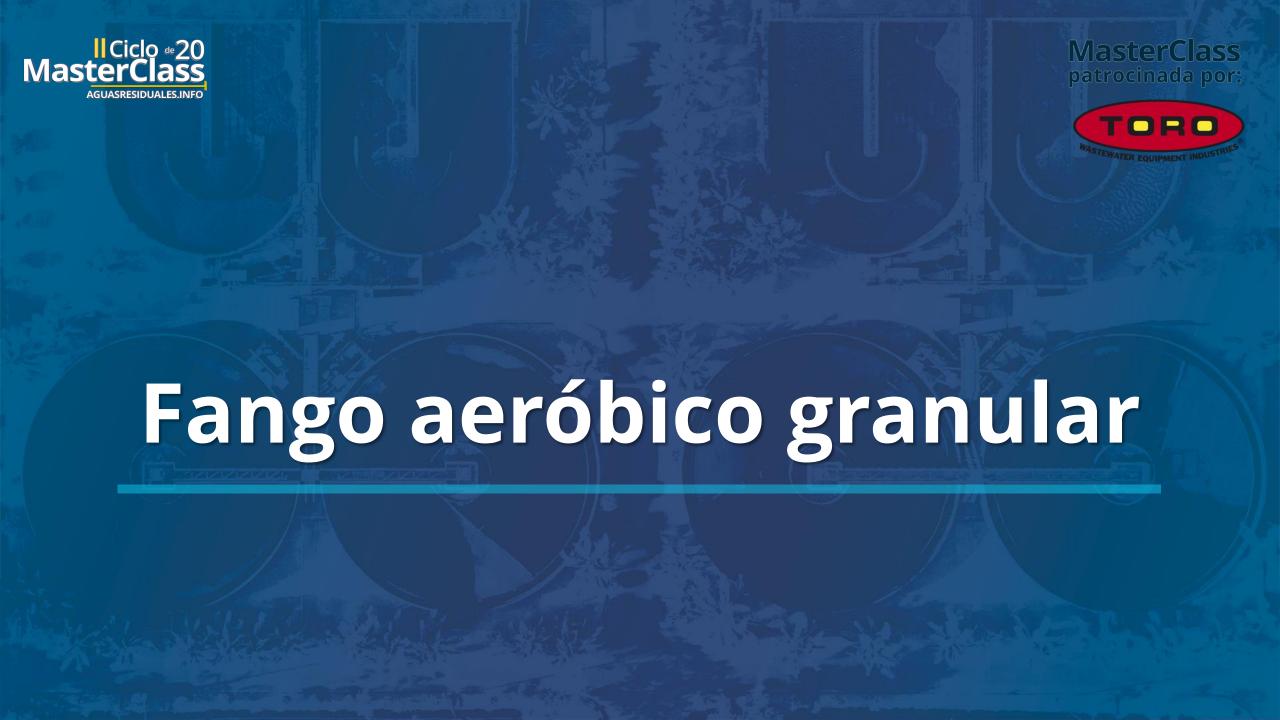
Fango aeróbico granular

Ventajas

Operación

Aplicaciones

Conclusiones

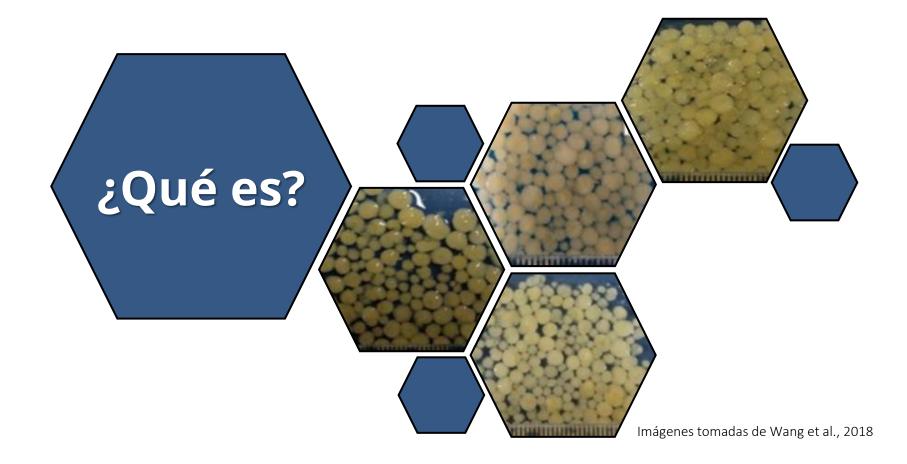




Fango aeróbico granular







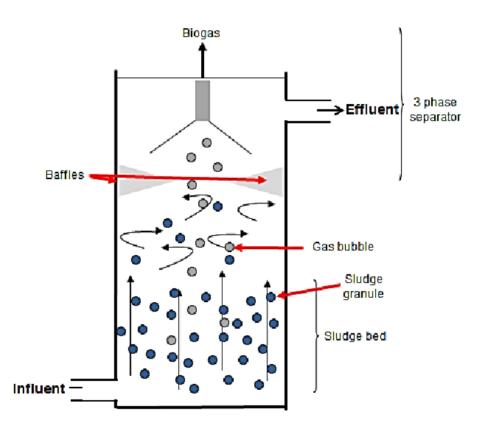


Fango granular





Lettinga et al., 1980



https://www.netsolwater.com/what-is-meant-by-the-uasb-reactor-its-working-and-quality-of-treated-water-using-a-uasb-reactor.php?blog=30



Fango granular



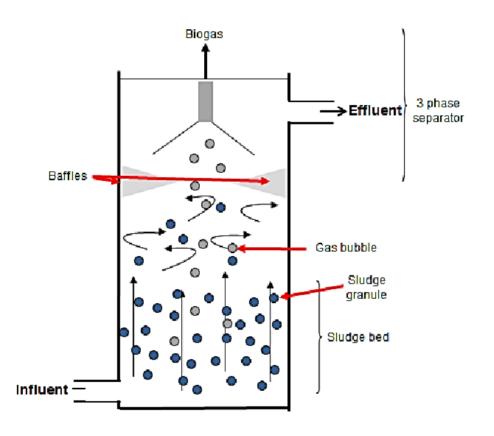


Lettinga et al., 1980

Beneficiado por

Bajo TRH

Alta carga orgánica



https://www.netsolwater.com/what-is-meant-by-the-uasb-reactor-its-working-and-quality-of-treated-water-using-a-uasb-reactor.php?blog=30



Fango granular





Lettinga et al., 1980

Beneficiado por

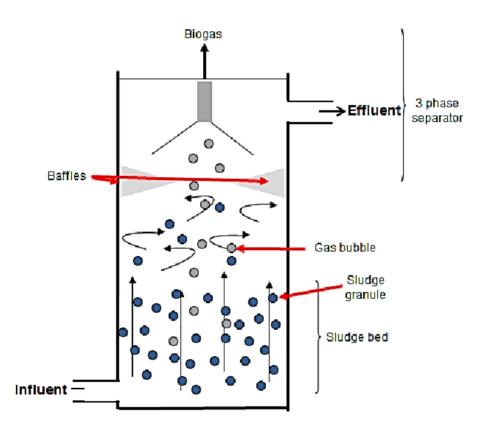
Bajo TRH

Alta carga orgánica

Formación

Movimiento

Sustancias poliméricas extracelulares



https://www.netsolwater.com/what-is-meant-by-the-uasb-reactor-its-working-and-quality-of-treated-water-using-a-uasb-reactor.php?blog=30



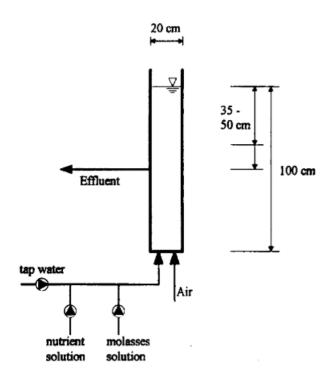
Fango aeróbico granular





Morgenroth et al., 1997





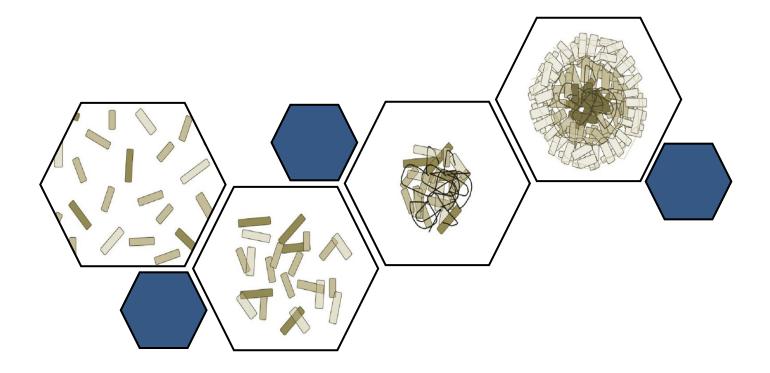
Morgenroth et al., 1997



Formación del gránulo





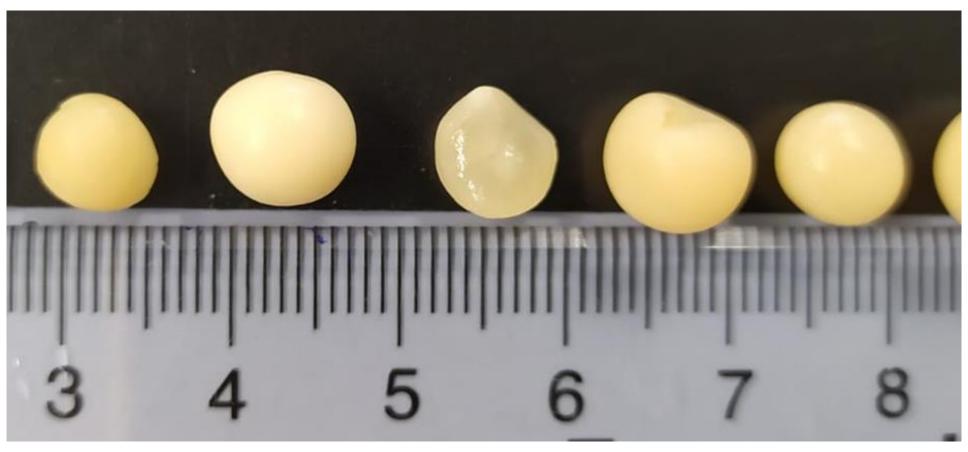


















Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente



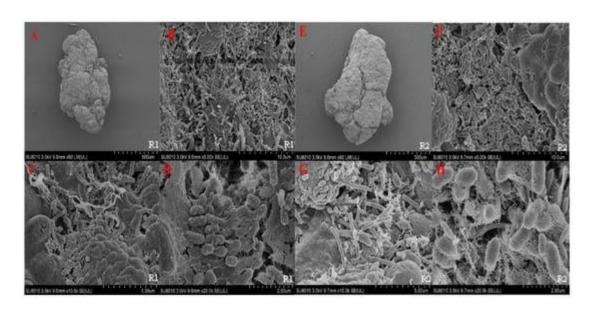




Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente



Zhang et al., 2022







Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente



Aqua-Aerobic Systems





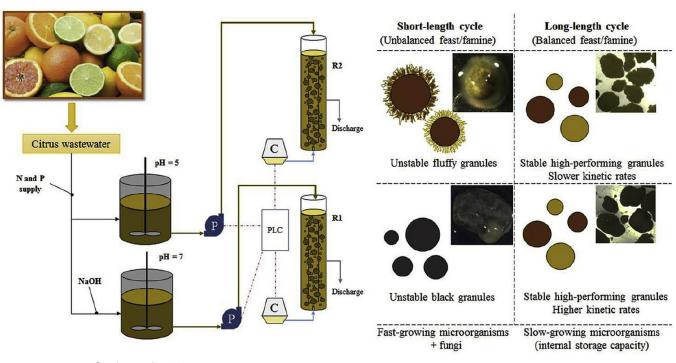


Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente

Eliminación simultánea de distintos contaminantes



Corsino et al., 2018



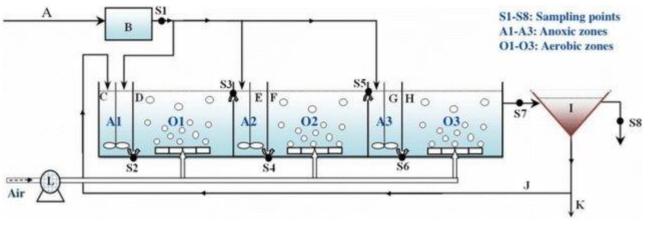




Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente



Chen et al., 2018



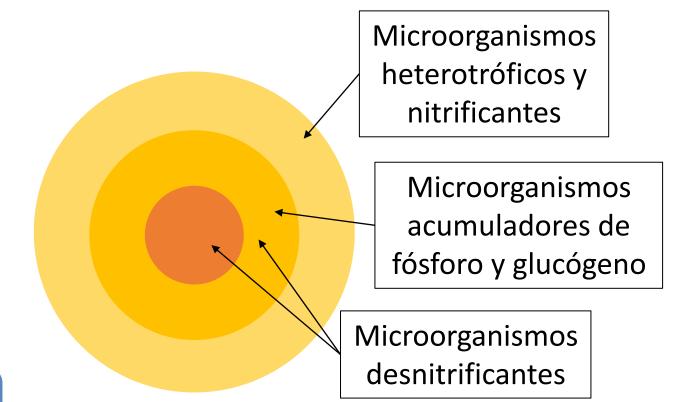




Mayor concentración de microorganismos

Rápida separación agua-biomasa

Resistencia a sustancias tóxicas y cambios en el influente



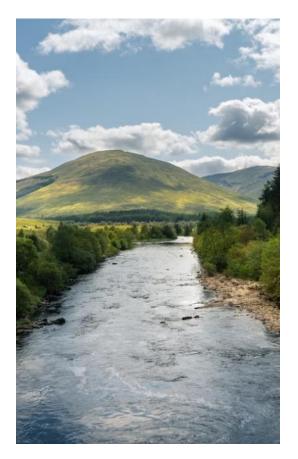


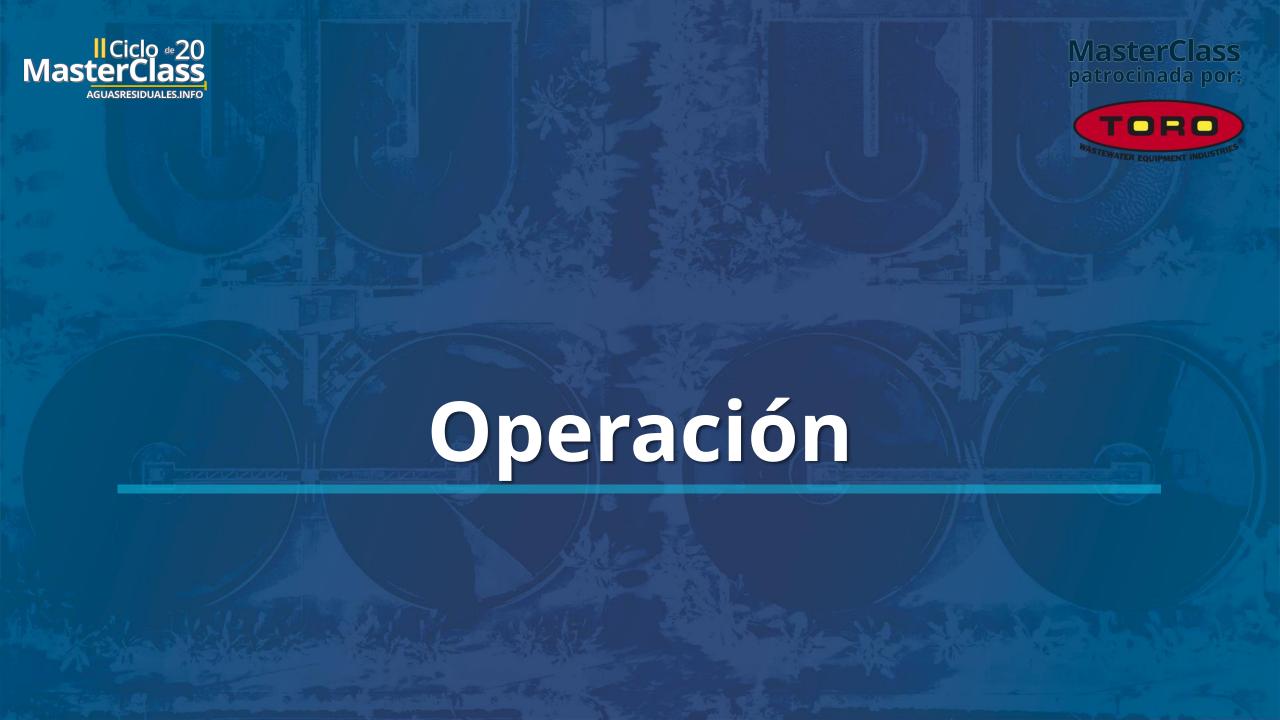








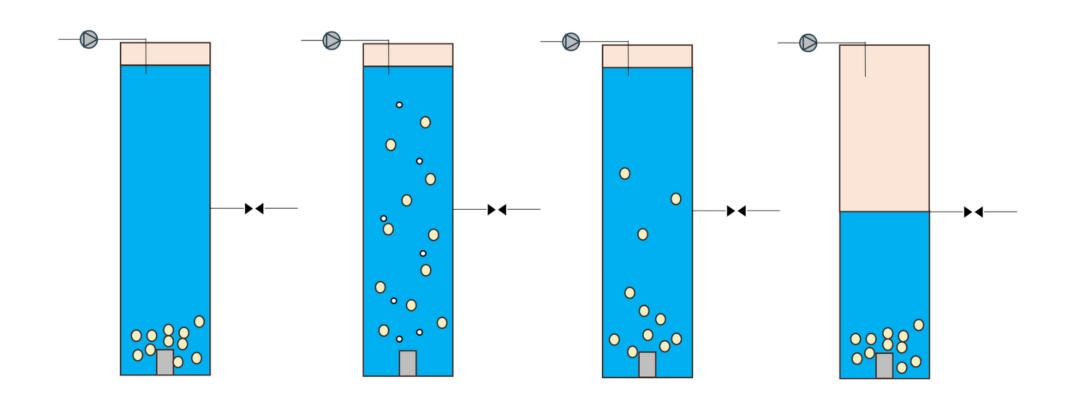








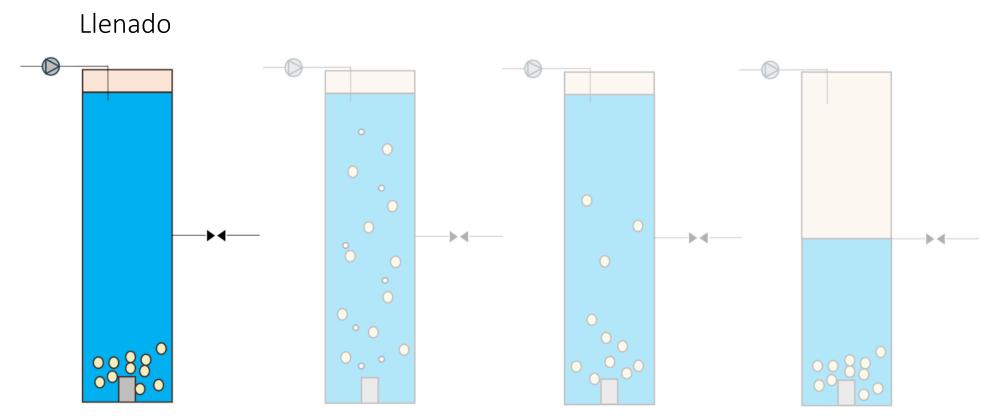










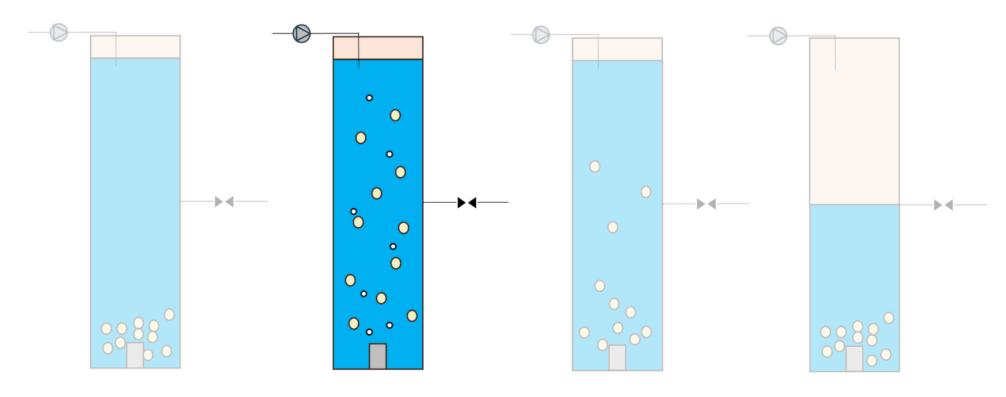








Aireación

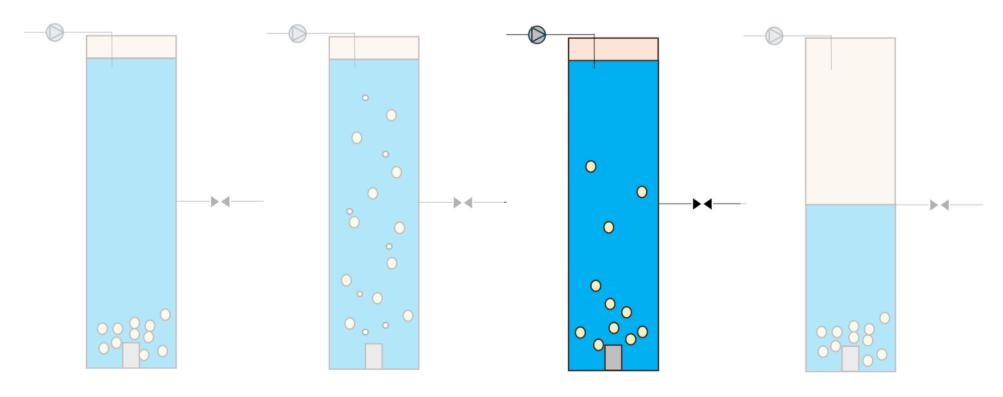








Decantación

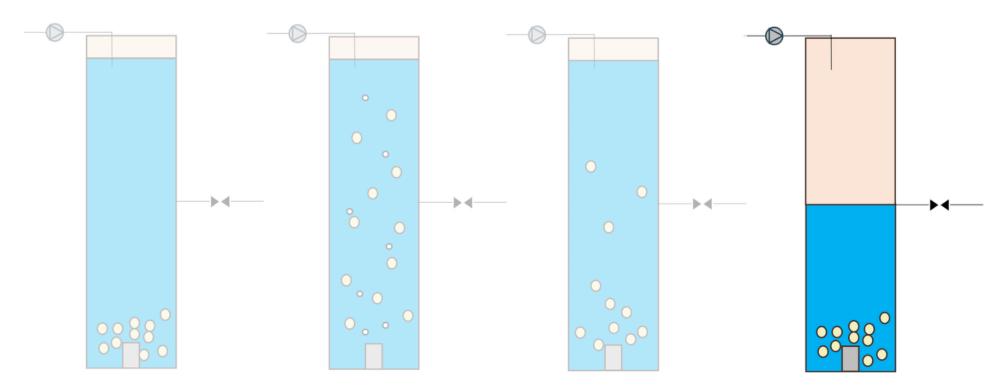








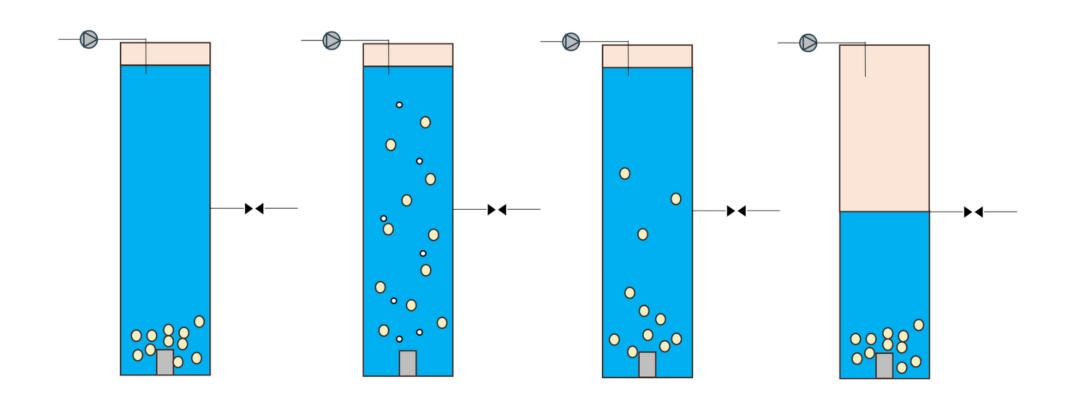
Vaciado









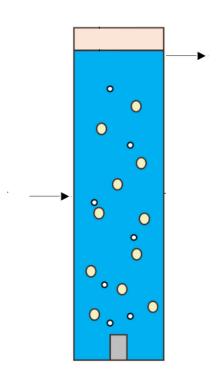




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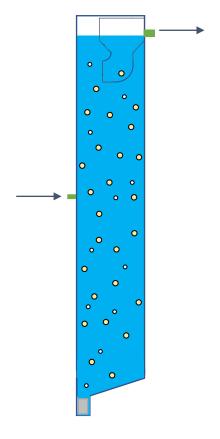








Continuo



MasterClass patrocinada por:



Journal of Environmental Chemical Engineering 11 (2023) 109938



Contents lists available at ScienceDirect Journal of Environmental Chemical Engineering







Description of new single-chamber continuous-flow reactors of aerobic granular sludge: Technical and biological study

Aurora Rosa-Masegosa a,b, Barbara Muñoz-Palazon b,c,*, Susanna Gorrasi c, Massimiliano Fenice^c, Alejandro Gonzalez-Martinez^{a,b}, Jesus Gonzalez-Lopez^{a,b}

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ARTICLE INFO Editor «Demo Batta-Kaminos»

Keywords: Aerobic granular sludg Granular stability microbial of

Aerobic granular studge reactors usually operate in sequential batch mode, although this configuration limits the treatment of large volumes of wastewater, and they require a storage system. To implement this technology at full-scale, it would be necessary to design a simple and compact continuous-flow bioreactor able to treat larger volumes of wastewater. In this study, four aerobic granular sludge single-chamber continuous-flow reactors (R1, R2, R3 and R4) were designed and operated at the lab scale to evaluate and select a bioreactor configuration that achieves high organic matter removal performance while maintaining a stable granulation for long-term operation, promoting the washout of filamentous microorganisms. Results confirmed that the bioreactor including a lateral decanter (R1), was able to work in a steady state without loss of granular biomass and reached 95% of organic matter removal performance. Its granules had excellent compaction, with settling velocity values above 100 m·h⁻¹. The R1 bioreactor also allowed rapid biomass adaptation and therefore a fast start-up (11 days).
Results of this preliminary study at the lab scale suggested that the new and simple bioreactor design could be promising for its implementation at full scale. For that, future research is required to optimise the current model and to determine the most suitable operational parameters to treat domestic wastewater at full scale.

Aerobic granular sludge (AGS) is a promising biological system for the treatment of wastewater due to its advantages in comparison with other technologies, such as conventional activated sludge (CAS) [1-3]. The hydrodynamic shear force and the continuous circular motion generate compact granules formed by microorganisms fixed and stabilised in a polymeric matrix [4,5]. This dense structure gives rise to the advantages of this technology because the high density of AGS promotes a better settleability [6,7], which can be translated into the implementation of more compact wastewater treatment plants (WWTPs) in comparison with CAS, due to lower time and space required to separate liquid-solid phases [2,5,8]. Moreover, the dense granular structure encourages a high accumulation of biomass. According to Nancharaiah et al. [9], AGS technology can reach a biomass concentration greater than 10 g-L⁻¹. Besides this, the round shape maximises the granular surface, and the high compactness promotes mass transfer, which creates differences in terms of oxygen and nutrients from the external to

internal layers [10-12]. This fact promotes the stratification of microorganisms along the layers, and consequently, it is possible to find different metabolisms in the same granule [2-4,13]. Therefore, AGS can remove organic matter, phosphorous, nitrogen and other substances, including pharmaceuticals, endocrine disrupters, phenolic compounds, dyes, heavy metals, particulate matter, nuclear waste and sulphur amino could modify influent characteristics [18]. However, AGS is a robust technology able to adapt to oscillating influent composition [19]. This capability is promoted by the large amount of extracellular polymeric substances (EPS) excreted by microorganisms in these aggregates, encouraging resistance to toxic compounds and high organic loading rates [13,20]. For all these reasons, AGS systems can be used to treat urban and industrial wastewater [13,21] as well as drinking water [22].

This technology has usually been operated in sequential batch reactors (SBRs), with the following cycles: 1) reactor filling with the raw water, 2) aeration, 3) settling of granules to separate the treated water from the biomass and 4) effluent discharge. In the last stage, light flocs

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Rosa-Masegosa et al., 2023

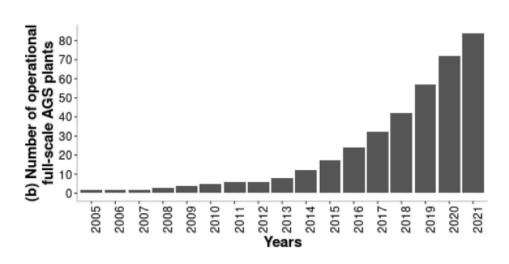
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 E-mail address: https://doi.org/10.1007/j.

















	Location	Operation start	Wastewater type
*	Ede, The Netherlands*	2005	Industrial - Diary
*	Rotterdam, The Netherlands*	2005	Industrial - Edible oil processing
	Gansbaai, South Africa*	2008	Municipal
*	Oosterwolde, The Netherlands*	2009	Industrial - Food processing
	Haining, China**	2010	30% domestic + 70%
			industrial
	Epe, The	2011	Municipal
	Netherlands* Vroomshoop, The Netherlands*	2013	Municipal
	Dinxperlo, The Netherlands*	2013	Municipal
	Garmerwolde, The Netherlands*	2014	Municipal
	Frielas, Lisbon*	2014	Municipal
	Switzerland:	2014	Municipal
	Utrecht, The Netherlands (pilot)*	2014-2021	Municipal
	Clonakilty, Ireland*	2015	Municipal
	Rykí, Poland*	2015	Municipal
	IJsselstein, The Netherlands*	2015	Industrial
	Wemmershoek, South Africa*	2015	Municipal
	Germany‡	2015	Municipal
	Carrigtwohill, Ireland*	2016	Municipal

Deodoro, Brazil*	2016	Municipal
Kingaroy, Australia*	2016	Municipal
Simpelveld, The Netherlands*	2016	Municipal
Deodoro, Rio de Janeiro, Brazil*	2016	Municipal
Sha Tin, Hong Kong *	2016	Municipal
Germany‡	2016/2017	Municipal
Rockford, USA*	2017	Municipal
Alpnach,	2017	Municipal
Switzerland*		_
Cork Lower Harbour, Ireland*	2017	Municipal
Jardim Novo, Rio Claro, Brazil*	2017	Municipal
Dublin Ringsend, Ireland*	2017	Municipal
Dublin - Ringsend, Ireland*	2017	Municipal
Germany‡	2017	Municipal
Israel‡	2017	Municipal
Faro – Olhão Portugal*	2018	Municipal
Great Dunmow, United Kingdom*	2018	Municipal
	2018	Municipal

	Morecambe, United Kingdom*	2018	Municipal
	Utrecht, The Netherlands*	2018	Municipal
k	Zutphen, The Netherlands*	2018	Industrial - Diary
	Lanaken, Belgium*	2018	Industrial – Fiber processing
	Strömstad, Sweden*	2018	Municipal
	Poland:	2018	Municipal
	Chile‡	2018	Municipal
	Dungannon, United Kingdom*	2019	Municipal
	Inverurie, United Kingdom*	2019	Municipal
	Radcliffe, United Kingdom*	2019	Municipal
	Walsall Wood, United Kingdom*	2019	Municipal
	Weert, The Netherlands*	2019	Municipal
	Wolf Creek Alabama, USA*	2019	Municipal
	Araguaína, Tocantis, Brazil*	2019	Municipal
	Tatu, Limeira, Brazil*	2019	Municipal
	Tijuco Preto, Sumaré, São Paulo. Brazil*	2019	Municipal

Paulo, Brazil* Quezon City, Philippines* Breskens, The Netherlands*	2019 2019	Municipal Municipal
Germany‡	2019	Municipal
Germany‡	2019	Industrial
Denmark‡	2019	Municipal
Denmark‡	2019	Municipal
Bangu, Brazil*	2020	Municipal
Dodewaard, The Netherlands*	2020	Municipal
Failsworth, United Kingdom*	2020	Municipal
Kloten – Opfikon, Switzerland*	2020	Municipal
Newham, United Kingdom*	2020	Municipal
Oissery, France*	2020	Municipal
Quakers Hill, Australia*	2020	Municipal
Saoneor, France*	2020	Municipal
Stein, The Netherlands*	2020	Municipal
Jardim São Paulo, Recife, Brazil*	2020	Municipal
Fleury, France*	2020	Municipal
Hartebeesfontein, South Africa*	2020	Municipal
Panheel, The Netherlands*	2020	Municipal
Switzerland‡	2020	Municipal

Fortaleza, Ceará, Brazil∴	2020	Municipal
Altena, Germany*	2021	Municipal
Longford, Australia*	2021	Municipal and industrial
Região Sul de Palmas, Brazil*	2021	Municipal
São Lourenço, Brazil*	2021	Municipal
Santo António, Goiánia, Brazil*	2021	Municipal
São Lourenço da Mata, Recife, Brazil*	2021	Municipal
Vriezenveen, The Netherlands*	2021	Municipal
Westnewton, United Kingdom*	2021 (Design)	Municipal
Wolcott Kansas,	2021 (Construction)	Municipal
Whitefish Montana, USA*	2021 (Construction)	Municipal
Norman Oklahoma, USA*	2021 (Design)	Municipal
Idaho Springs Colorado*	2021 (Start-up pending)	Municipal
Jaboatão, Recife, Brazil*	2022	Municipal
Jeddah, Saudi Arabia*	2023	Municipal
Muskiz, Spain*	2023	NA
Camaragibe, Brazil*	2025 (Design)	Municipal



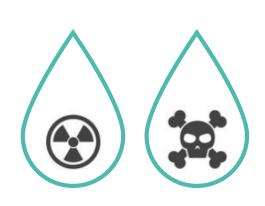
Industrial

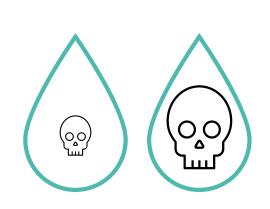


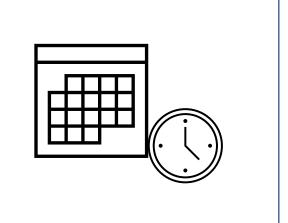














Industrial



















- Materia orgánica
- N, P, S...
- Sólidos en suspensión
- Salinidad
- pH
- Pesticidas
- Grasas
- Compuestos fenólicos
- **Detergentes**
- Patógenos





Industrial: Almazara





Contaminantes

- Materia orgánica
- Grasas
- Sólidos en suspensión
- Compuestos fenólicos
- Sales (conductividad)
- pH ácido





Industrial: Almazara (ácidos fenólicos)





Journal of Hazardous Materials 376 (2019) 58-67

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Journal of Hazardous Materials





Performance and microbial community structure of an aerobic granular sludge system at different phenolic acid concentrations



Barbara Muñoz-Palazon^{*}, Alejandro Rodriguez-Sanchez, Miguel Hurtado-Martinez, Ines Manuel de Castro, Belén Juarez-Jimenez, Alejandro Gonzalez-Martinez, Jesus Gonzalez-Lopez

Institute of Water Research, University of Granada, C/Ramon y Cajal, 4, 18071, Granada, Spain

LEINFO

Keywords: Phenolic compounds Aerobic granulur sludge Tooic compounds Olive washing water (OWW) ABSTRAC

The present work aims to use aerobic granular sludge technology for the treatment of wastewater comining high organic matther loads and a mixture of phenolic compounds normally persent in olive weaking water. The physicochemical performance of five bioractors treating different concentrations of mixture of phenolic acid was monitored to observe the response of the systems. The bioexectors trutted persented at 50, 100 and 300 mg L⁻¹ did not show relevant changes in terms of performance and granular perspective, showing high ratio of phenolic compound removal ratio. However, the bioractors operated with high phenolic compound concentrations showed low rates of organic matter, nitrogen and phenolic acid removal. In the same way, high concentrations showed low rates of organic matter, nitrogen and phenolic acid removal. In the same way, high concentrations of phenolic compound eleramented the dismingration of the granular bloams. Next generation sequencing studies showed a stable community structure in the bioractors operating with 50, 100 and 500 mg. L⁻¹ of phenolic distributions promoted dominant phyloripes. Conversely, the reactors operated at 180 and 600 mg. L⁻¹ of phenolic distributions promoted the profiferation of Oligothymanophorus citiates. Thus, this study suggests that aerobic granular studge technology could be used fifther the treatment of vastressers contact and colive washing water.

1. Introduction

Phenolic compounds are chemical substances that make important contributions to the pollutant and recalcitrant character of different industrial and agricultural wastewaters [1,2]. Different industrial and agri-food effluents present high concentrations of phenolic acids, those of the olive oil industry being of particular importance in the Mediterranean area, with more than 3 million tons produced annually mainly by Spain, Portugal, Italy and Greece [6]. The main effluent generated during the production of olive oil is olive washing water (OWW) [7-9]. Regarding olive oil by-products, phenolic compounds are quantitatively abundant [10], with total phenol content between 200 and 300 mg L-1. The main families of phenolic compounds identified by authors in olive mill wastes are phenolic acids, secoiridoids and flavonoids [10,11]. Moreover, OWW presents a high organic matter load, with a chemical oxygen demand (COD) of up to 5 g L-1. Moreover, phenolic compounds deserve special attention due to their influence as antibacterial and phytotoxic products [3], even at very low concentrations, limiting the use of biological processes. In this context,

advanced oxidation, membrane filtration, biological digestion, heterogenous photocatalysis and Photo-Fenton degradation have been used for phenolic acid degradation [1,4], although these technologies have serious disadvantages such as their economic cost and not being very friendly to the environment. It is for this reason that the search for biological alternatives for the treatment of effluents polluted with phenolic compounds is a challenge of great biotechnological interest. Thus, in the last decade, aerobic-biodegradation processes such as submerged biofilter or photobioreactor(PBR) systems have been proposed due to the robustness for the treatment of agricultural waste waters polluted with high concentrations of phenolic substances [1,5] The PBR was used for treating real olive wastewater, and the results confirmed that the biofilm formed showed a high capacity to remove phenols, nitrogen, phosphorous and organic matter, also the system tention time. On the other hand, it was proposed a innovative biological technology denominated aerobic granular sludge technology for treating phenols and pesticides, between among pollutants, which are

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Muñoz-Palazón et al., 2019

Reactor: SBR, escala laboratorio.

TRH: 6 h.

Inóculo: gránulos.

Mezcla de Ácido caféico, ácido hidroxibenzoico, ácido

protocatecuico. 50, 100, 300, 600 y 1000 mg/L.



Industrial: Almazara (ácidos fenólicos)





Journal of Hazardous Materials 376 (2019) 58-67

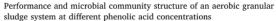


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Barbara Muñoz-Palazon°, Alejandro Rodriguez-Sanchez, Miguel Hurtado-Martinez, Ines Manuel de Castro, Belén Juarez-Jimenez, Alejandro Gonzalez-Martinez, Jesus Gonzalez-Lopez

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1. Introductio

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Muñoz-Palazón et al., 2019

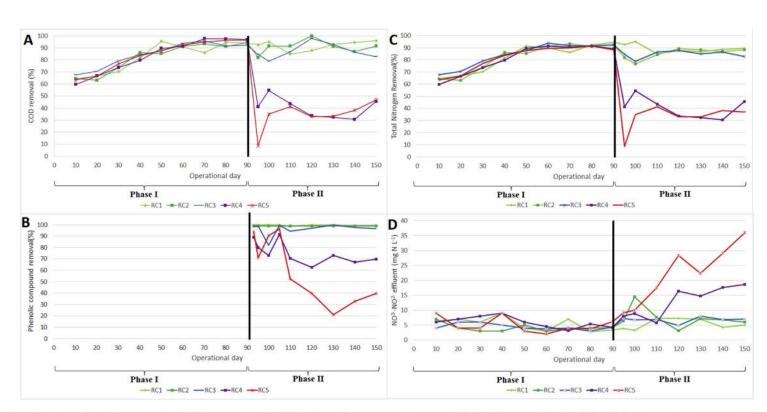


Fig. 2. Physic-chemical performance for COD removal ratio (A); phenolic compounds removal (B); TN removal ratio (C); nitrite and nitrate concentration in the effluent (D).



Industrial: Almazara (ácidos fenólicos)

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Institute of Water Research, University of Granada, C/Ramon y Caial, 4, 18071, Granada, Spain

ARTICLE INFO

Keywords: Phenolic compounds Aerobic granular sludge Olive washing water (OWW) The present work aims to use aerobic granular sludge technology for the treatment of wastewater containing high organic matter loads and a mixture of phenolic compounds normally present in olive washing water. The physicochemical performance of five bioreactors treating different concentrations of mixture of phenolic acid was monitored to observe the response of the systems. The bioreactors that operated at 50, 100 and 300 mg L did not show relevant changes in terms of performance and granules properties, showing high ratio of phenolic compound removal ratio. However, the bioreactors operated with high phenolic compound concentration showed low rates of organic matter, nitrogen and phenolic acid removal. In the same way, high concentrations of phenolic compounds determined the disintegration of the granular biomass. Next-gene showed a stable community structure in the bioreactors operating with 50, 100 and 300 mg L-1 of phenolic acids, with the genera Lampropedia and Arenimonas, family Xanthobacteraceae and Fungi Pezizomycotina as the

dominant phylotypes. Conversely, the reactors operated at 500 and 600 mg L⁻¹ of phenolic substances promoted proliferation of Oligohymenophorea ciliates. Thus, this study suggests that aerobic granular sludge technology could be useful for the treatment of wastewaters such as olive washing water.

1. Introduction

Phenolic compounds are chemical substances that make important

genous photocatalysis and Photo-Fenton degradation have been used for phenolic acid degradation [1,4], although these technologies have erious disadvantages such as their economic cost and not being very friendly to the environment. It is for this reason that the search for biological alternatives for the treatment of effluents polluted with phenolic compounds is a challenge of great biotechnological interest. Thus, in the last decade, aerobic-biodegradation processes such as submerged biofilter or photobioreactor(PBR) systems have been proposed due to the robustness for the treatment of agricultural waste waters polluted with high concentrations of phenolic substances [1,5] The PBR was used for treating real olive wastewater, and the results confirmed that the biofilm formed showed a high capacity to remove phenols, nitrogen, phosphorous and organic matter, also the system reached the steady-stable in a short period under 2 h of hydraulic retention time. On the other hand, it was proposed a innovative biological technology denominated aerobic granular sludge technology for treating phenols and pesticides, between among pollutants, which are

advanced oxidation, membrane filtration, biological digestion, betero

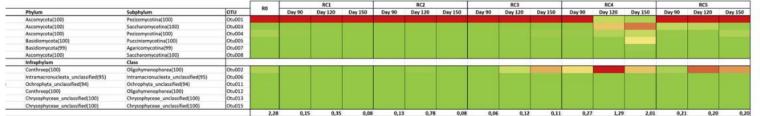
contributions to the pollutant and recalcitrant character of different industrial and agricultural wastewaters [1,2]. Different industrial and agri-food effluents present high concentrations of phenolic acids, those of the olive oil industry being of particular importance in the Mediterranean area, with more than 3 million tons produced annually mainly by Spain, Portugal, Italy and Greece [6]. The main effluent generated during the production of olive oil is olive washing water (OWW) [7-9]. Regarding olive oil by-products, phenolic compounds are quantitatively abundant [10], with total phenol content between 200 and 300 mg L-1. The main families of phenolic compounds identified by authors in olive mill wastes are phenolic acids, secoiridoids and flavonoids [10,11]. Moreover, OWW presents a high organic matte load, with a chemical oxygen demand (COD) of up to 5 g L-1. More over, phenolic compounds deserve special attention due to their influence as antibacterial and phytotoxic products [3], even at very low concentrations, limiting the use of biological processes. In this context,

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Muñoz-Palazón et al., 2019







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Contaminantes

- Materia orgánica
- Aceites esenciales
- pH < 3
- Sólidos en suspensión
- Ausencia de N y P





Reactor: SBR, escala laboratorio.

TRH: 24 y 12 h.

Inóculo: fango granular

pH 7 y pH 5

Composition of the feeding wastewater in R1 and R2.

Parameter	Unit measure	Raw citrus wastewater	Period I (30 days)	Period I (30 days)		Period II (30 days)	
			R1	R2	R1	R2	
Total COD (TCOD)	mg L ⁻¹	3065-7512 (5464 ± 1291)	3065-7300 (5579 ± 1434)	3065-7300 (5579 ± 1434)	3175-7512 (5296 ± 1319)	3175-7512 (5296 ± 1319)	
Soluble COD (SCOD)	mg L ⁻¹	1595-6480 (3823 ± 1473)	1595-6045 (4029 ± 1349)	1595-6045 (4029 ± 1349)	1920-6480 (3525 ± 1615)	1920-6480 (3525 ± 1615)	
Total Nitrogen (TN)	mg L ⁻¹	12 ± 3.5	40-151 (101 ± 13)	40-151 (101 ± 13)	48-162 (88 ± 9)	48-162 (88±9)	
Total Phosphorous (TP)	mg L ⁻¹	3.5 ± 1.1	8-30 (20 ± 3)	8-30 (20 ± 3)	9.3-32.4 (17.6 ± 2)	9.3-32.4 (17.6 ± 2)	
Total suspended solids (TSS)	mg L ⁻¹	300-1914 (1219 ± 652)	300-1604 (906 ± 321)	300-1604 (906 ± 321)	1004-1914 (1324 ± 426)	1004-1914 (1324 ± 426)	
pH	_	5.2 ± 0.2	7.2 ± 0.3	5.3 ± 0.2	7.4 ± 0.4	5.1 ± 0.2	

Legend: min-max (average ± std.dev) values.

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Research article

Aerobic granular sludge treating high strength citrus wastewater: Analysis of pH and organic loading rate effect on kinetics, performance and stability



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ARTICLEINFO

Article history: Received 4 September 2017 Received in revised form 25 January 2018 Accepted 24 February 2018 Available online 5 March 2018

Keywords: Aerobic granular sludge Biokinetics Citrus wastewater pH

ABSTRACT

In the present paper, the feasibility of citrus wastewater treatment with aerobic granular sludge sequencing bather neators (ACSB) was investigated. Two ACSBRS (Damed R1 and R2, sepectively) were operated for 90 days under different organic loading rates (OR1) and pH in two experimental periods. The OLK ranged approximately between 3 okg TCOD m³d⁻¹ and 78 pCTOD m³d⁻¹ and 78 pCTOD m³d⁻¹ and 15 kg TCOD m³d

The results revealed that under high OLR and unbalanced feast/famine regime (Period I), the development of fast-growing microorganisms (fungi and filamentous bacteria) was favoured in both reactors, resulting in granular shadge instability. An extended famine phase and a proper balancing between feast and famine periods (Period II) were favourable for the development of bacteria with low growth rates (10.05 d⁻¹) thus enhancing the granulest stability. To the benefit of granular sludge stability and effluent quality, the length of the feast period should not exceed 25% of cycle length.

Moreover, under OLR lower than 75x (TOOI) m⁻²d⁻¹ the removal efficiency of total chemical oxygen

Moreover, under O.R. lower than 7 kg TCOD m."d" "the removal efficiency of total chemical oxygen demand (TCOD) was approximately 950 in R1 and R2 and no side effects on the organic carbon removal performance related to the pH were observed. In contrast, at higher OLR a significant decrease in the removal efficiency (from 90% to less than 755) was observed in R2. Results revealed also that under low pH, hydrohysis of proteins occurred and a decrease in the biological kinetic rates proportionally to the anoiled OLR was observed.

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1. Introduction

The citrus industry requires large amounts of water for washing of fruits and equipment, Juices and essential oils extraction, etc. Citrus is largely cultivated in southern Italy (Zema et al., 2012), especially in Sicily. It was estimated that a citrus factory that process 25 th of lemon produces over 10 million litres of wastewater per day (Navarro et al., 2008). Citrus wastewater is highly contaminated with organic matter, suspended solids, high content of essential oils and acidity (pH - 3), as well as unbalanced nutrients due to the lack in nitrogen and phosphorous (Zema et al., 2012). Moreover, citrus wastewater is characterized by seasonal qualitative and quantitative variability (El-Kamah et al., 2010).

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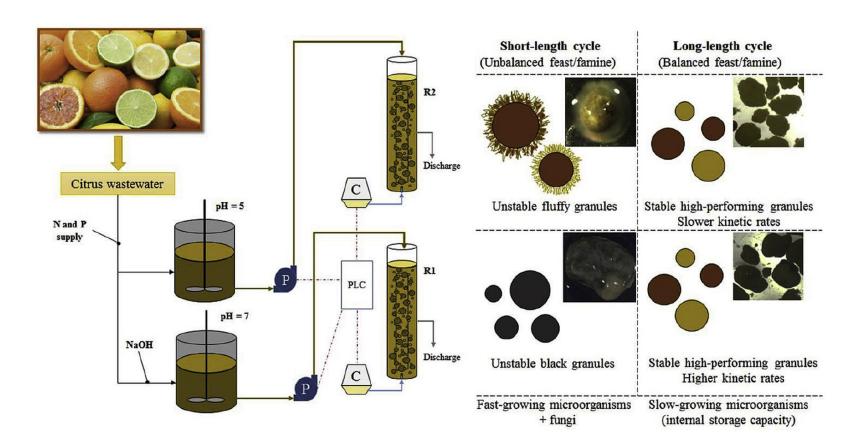
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Corsino et al., 2018

Indeed, chemical oxygen demand (COD) concentration varies greatly according to the different stages of the production process. In Sicily, the highest peak of production is generally reached toward to beginning of March, because of the simultaneous maximum productivity in oranges, lemons and mandarins. Furthermore, the qualitative composition of citrus wastewater could change through the workday, basing on the production step (as juice or essential oil extraction, or fruit washing, etc.). Essential oils are responsible for their antimicrobial actions, hence their presence, although in low concentration, could be a limit for biological processes (Saverini et al., 2012; Zema et al., 2012). Overall, the composition and the extreme variability of the wastewater quality, in terms of organic loading and alikalinity, represent a serious challenge for biological treatments.

Ensuring compliance with stringent regulatory standards for discharge into a municipal sewage system or receiving water bodies requires highly efficient technologies to manage properly





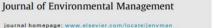
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Research article

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ABSTRACT

In the present paper, the feasibility of citrus wastewater treatment with aerobic granular sludge sequencing batch reactors (ACSBQ) was investigated. Two ACCSBS (named RI and RZ, respectively) were operated for 90 days under different organic loading rates (OLR) and pit in two experimental periods. The OLR ranged approximately between 3 blg TCOO m ^{3-d} and r/s R TCOO m ^{3-d} thring Ferold, whereas between 7 kg TCOO m ^{3-d} - 1 and 15 kg TCOO m ^{3-d} - 1 during Period II, pH was maintained at 70 and 5.5 in R1 and R2, respectively.

The results revealed that under high OLR and unbalanced feast/famine regime (Period I), the development of fast-growing microorganisms (fungi and filamentous bacteria) was favoured in both reactors, resulting in granular studge instability. An extended famine phase and a proper balancing between feast and famine periods (Period II) were favourable for the development of bacteria with low growth raster (10.05 d⁻¹) thus enhancing the granulest stability. To the benefit of granular studge stability and efflicent quality, the length of the feast period should not exceed 25% of cycle length.

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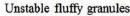
Corsino et al., 2018

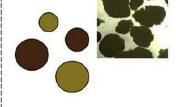


<u>DQO %</u> <u>eliminación</u>	TRH 12 h	TRH 24 h	
pH ácido	75 %	90 %	
pH neutro	90 %	90 %	

Short-length cycle (Unbalanced feast/famine) Long-length cycle (Balanced feast/famine)







Stable high-performing granules Slower kinetic rates



Unstable black granules

Fast-growing microorganisms + fungi



Stable high-performing granules Higher kinetic rates

Slow-growing microorganisms (internal storage capacity)

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Research article

Aerobic granular sludge treating high strength citrus wastewater: Analysis of pH and organic loading rate effect on kinetics, performance and stability



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ABSTRACT

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Corsino et al., 2018







Contaminantes

- Materia orgánica
- Sólidos en suspensión
- pH ácido
- Ausencia de nutrientes





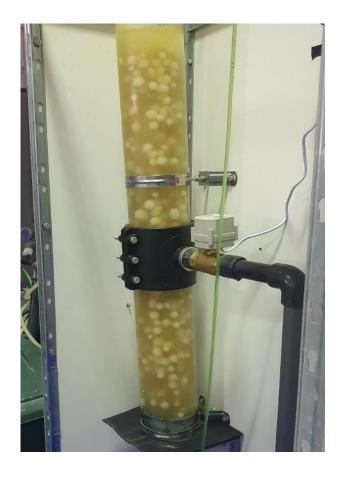




Reactor: SBR, escala laboratorio.

TRH: 6 h.

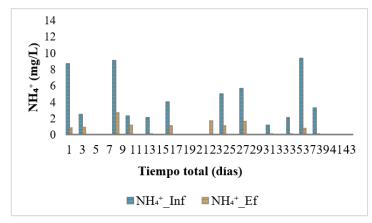
Pretratamiento: neutralización

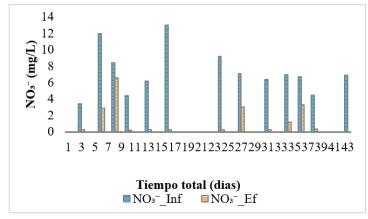


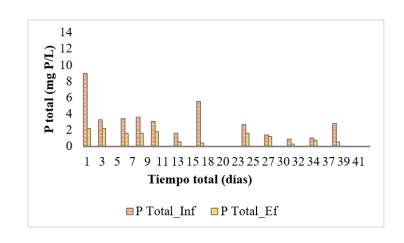








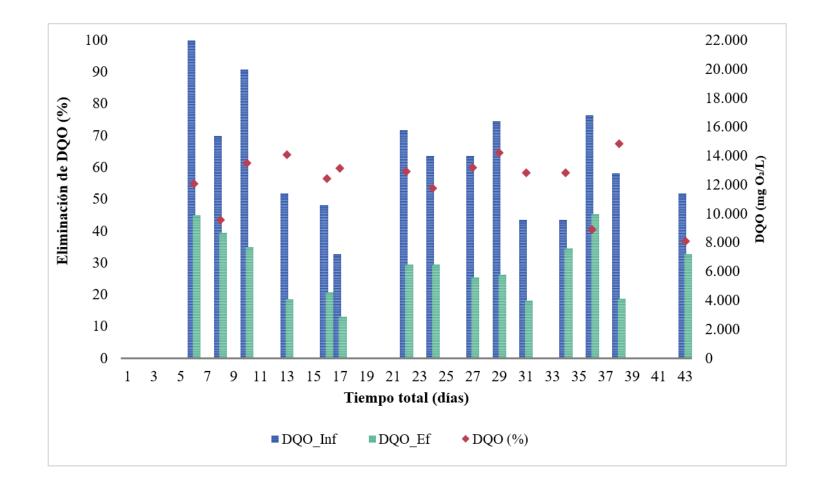








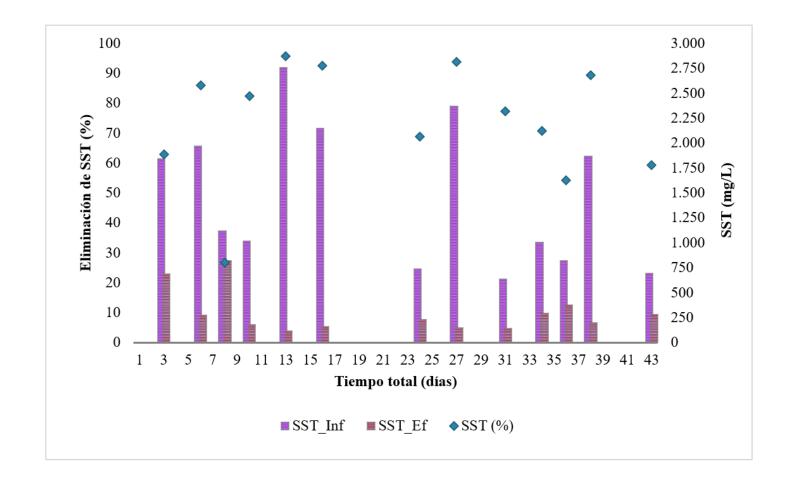










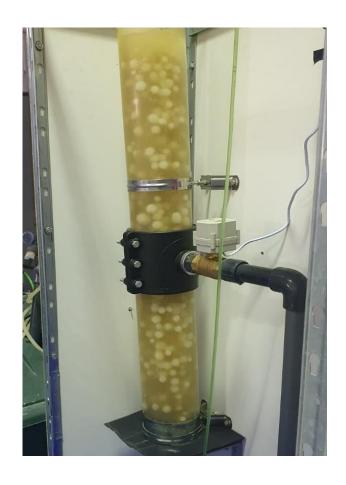


















Contaminantes

- Materia orgánica
- Grasa
- N, P
- Sólidos en suspensión





Reactor: SBR, escala laboratorio.

TRH: 8 h.

Inóculo: fango flocular

Average characteristics of the influent wastewater

Parameter (units)	Measured value
pH	7.3±0.4 (29) ^a
Total COD (mg l ⁻¹)	$7685 \pm 646 (24)$
SCOD (mg l ⁻¹)	$5163 \pm 470 (24)$
$TKN (mg l^{-1})$	$1057 \pm 63 (24)$
$NH_4-N \ (mg \ l^{-1})$	$50 \pm 12 (24)$
$NO_3-N \text{ (mg l}^{-1})$	0 ± 0 (24)
$NO_2-N \ (mg \ l^{-1})$	$0 \pm 0 (24)$
Total P (mg l ⁻¹)	217 ± 38 (24)
Alkalinity (mg1 ⁻¹ as CaCO ₃)	$872 \pm 42 (8)$
SS $(mg l^{-1})$	1742 ± 116 (23)
VSS (mg l ⁻¹)	$1520 \pm 128 (23)$

^aMean ± standard deviation (number of measurements).









Water Research 39 (2005) 4817-4823

Nitrogen and phosphorus removal from an abattoir wastewater in a SBR with aerobic granular sludge

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bstract

The formation and performance of granular sludge was studied in an 81 sequencing batch reactor (SBR) treating an abattoir (slaughterhouse) wastewater. Influent concentrations averaged 1520 mg l⁻¹ volatile suspended solids (VSS), 7685 mg l⁻¹ Chemical oxygen demand (COD), 1057 mg l⁻¹ total kjeldahl nitrogen (TKN), 217 mg l⁻¹ total P. The COD loading was 2.6 kg m⁻³ d⁻¹. The SBR was seeded with flocculating sludge from a SBR with an 1h settle time, but granules developed within 4 days by reducing the settle time to 2 min. The SBR cycle also had 120 min mixed (anaerobic) fill, 220 min aerated react, and 18 min draw/idle. The granules had a mean diameter of 1.7 mm, a specific gravity of 1.035, a density of 62 g VSS1⁻¹, a zone settling velocity (ZSV) of 51 m l⁻¹, and a sludge volume index (SVI) of 22 mlg l⁻¹. Without optimizing process conditions, removal of COD and P were over 98%, and removal of N and VSS were over 97%. Nitrification and denitrification occurred simultaneously during react. The results indicate that conventional SBRs treating wastewaters with flocculating sludge can be converted to granular SBRs by reducing the settle time.

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Keywords: Abattoir; Aerobic granules; Biological phosphorus removal; Denitrification; Nitrification; SBR; Slaughterhouse; Wastewater

1. Introduction

Granular sludge was first described in strictly anaerobic systems, such as upflow anaerobic sludge blanket (UASB) reactors (Lettinga et al., 1980), biofilm airlift reactors (van Loosdrecht et al., 1995), and anaerobic sequencing batch reactors (SBRs) (Wirtz and Dague, 1996). Granules have a greater density and

diameter than floes, allowing a bioreactor to maintain high biomass concentrations with excellent settling properties (van Loosdrecht et al., 1995). The resulting increase in reactor efficiency explains why over 900 UASB reactors have been built around the world (Alves et al. 2000).

Recently, the formation of granules under aerobic and alternating aerobic/anaerobic conditions has been reported in SBRs seeded with floculating sludge. Several factors promote the formation of aerobic granules from flocs. A short settle time is required to wash out freeliving and floc-forming microorganisms and apply selective pressure for granule formation (Morgenroth

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Cassidy & Belia, 2005

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Average effluent measurements and performance calculations for the SBR during steady state operation (day 16-day 76)

Parameter (units)	Measured value
рН	7.1 ± 0.3 (64) ^a
Total COD (mg l ⁻¹)	$106 \pm 18 (64)$
SCOD (mg1 ⁻¹)	$95 \pm 17 (64)$
Total COD removal (%)	98.6
$TKN (mg l^{-1})$	$2 \pm 2 (64)$
$NH_4-N \text{ (mg l}^{-1})$	0 (64)
$NO_3-N \ (mg \ l^{-1})$	26 (64)
$NO_2-N \text{ (mg l}^{-1})$	0 (64)
N Removal (%)	97.4
Total P (mg l ⁻¹)	4 ± 2 (64)
P Removal (%)	98.2
P Content of biomass (% by weight)	5.7 ± 0.8 (31)
EVSS (mg l ⁻¹)	$42 \pm 8 (64)$
VSS removal (%)	97.2

^aMean ± standard deviation (number of measurements).

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Water Research 39 (2005) 4817-4823

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Reactor: SBR, escala laboratorio, 20 L.

TRH: 12 h.

Inóculo: fango activo

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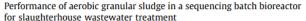


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Bioresource Technology



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HIGHLIGHTS

- · Biological nutrient removal was positive correlated to sludge granulation
- . The external feature of granular sludge was studied by 3D-EEM and EPS.
- Particle size distribution showed the superior size for AOB and NOB growth. . The SOUR was used to analyze the activity of AOB and NOB.

ARTICLE INFO Article history:

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Aerobic granular sludge Slaughterhouse wastewate Particle size Extracellular polymeric substances Specific oxygen demand rates

Lab-scale experiment was conducted to investigate the formation and characteristics of aerobic granular sludge for biological nutrient removal of slaughterhouse wastewater. Experimental results showed that removal performances of chemical oxygen demand (COD), ammonia and phosphate were enhanced with sludge granulation, and their removal efficiencies reached 95.1%, 99.3% and 83.5%, respectively. The aerobic granular sludge was matured after 90 days cultivation, and protein-like substances were the main components. Simultaneously, the mass ratio of proteins and polysaccharides (PN/PS) was enhanced to 2.5 from 1.7. The granules with particle sizes of 0.6-1.2 and 1.2-1.8 mm, accounting for 69.6%, were benefit for the growth of ammonia oxidizing bacteria (AOB) and nitrate oxidizing bacteria (NOB), and corresponding specific oxygen demand rates (SOUR) of AOB and NOB were 31.4 and 23.3 mgO2/

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1. Introduction

Slaughterhouse wastewater is characterized by high concentrations of organic matters, suspended solids, oil and grease, nitrogen and phosphate (Merzouki et al., 2005). Due to the high concentrations of organic matters and nitrogen, biological treatment of this wastewater is usually accomplished in units sequentially arranged to remove organic matter prior to nitrogen. Anaerobic processes normally remove a significant fraction of organic matters, but low nitrogen. Conventional nitrogen removal systems involve the installation of several units of sequential operations, requiring large areas for the deployment of full scale systems. One way to reduce the deployment areas is by systems that integrate nitrification and denitrification within a single unit (Bassin et al. 2012) After anaerobic hydrolysis and/or physical pretreatment of this

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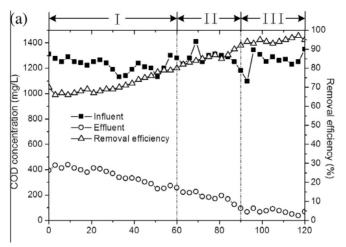
wastewater, a nitrification/denitrification system based on the aerobic granular technology can fulfill the environmental requirements (Zhao et al., 2013).

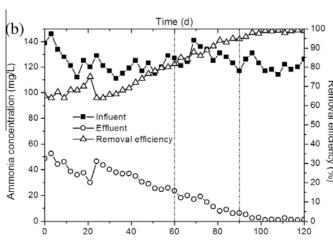
Aerobic granular sludge has attracted increasing attention in recent years, for its high metabolic activity, large granule diameter, remarkable settle ability, and high biomass retention without any support media, ability to withstand shock loadings and unfavorable environmental conditions, and especially, the excellent simultaneous nitrification and denitrification capacity (Adav et al., 2009; Wei et al., 2012). However, reports of successful aerobic granulation system were mainly focused on readily biodegradable synthetic wastewater (Nancharaiah and Venugopalan, 2011). Information on application of aerobic granules for the treatment of refractory wastewaters with high-strength ammonium, like phenol wastewater (Aday et al., 2007), agro-based wastewater (Abdullah et al., 2011), livestock wastewater (Othman et al. 2013), rubber wastewater (Rosman et al., 2014) and domestic wastewater (Coma et al., 2012) is still very initiated, even though biological treatment of these wastewaters has been practiced for

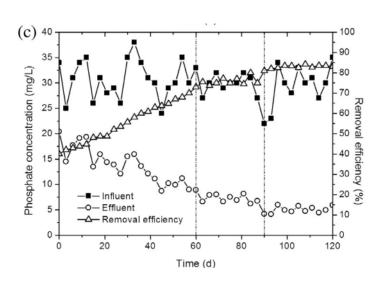
Liu et al., 2015

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Performance of aerobic granular sludge in a sequencing batch bioreactor



for slaughterhouse wastewater treatment Yali Liu a, Xiaorong Kang a,b,*, Xin Li a, Yixing Yuan

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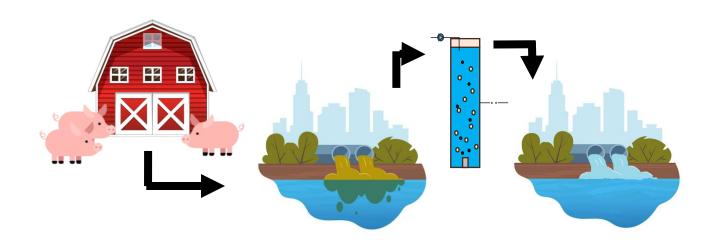
Industrial: Granjas





Contaminantes

- Materia orgánica
- N, P
- Sólidos en suspensión
- Aminoácidos
- Antibióticos





Industrial: Granjas (N y P)

Reactor: SBR, escala laboratorio.

TRH: 8 h.

Inóculo: mezcla de fango activo y lodo de la granja.

Agua: orina, sangre, heces y agua de lavado (tamiz 1 mm)

Parameters ^a	Livestock wastewater (cattle farm)		
рН	8.05		
Chemical oxygen demand (COD)	3600		
Biochemical oxygen demand (BOD)	1750		
Total suspended solid	230		
Total nitrogen	650		
Total phosporus	380		

All other parameters are in mgL⁻¹ except pH.

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Bioresource Technology 133 (2013) 630-634

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Short Communication

Livestock wastewater treatment using aerobic granular sludge



Inawati Othman ^a, Aznah Nor Anuar ^{b,*}, Zaini Ujang ^b, Noor Hasyimah Rosman ^a, Hasnida Harun ^a, Shreeshivadasan Chelliapan b

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- ▶ Matured aerobic granules were observed in the reactor with the size up to 4.1 mm
- Removal efficiency of 74% COD 73% TN and 70% TP can be achieved at 4-b cycle time

ARTICLE INFO

Received 19 November 2012 Received in revised form 21 January 2013 Accepted 29 January 2013 Available online 9 February 2013

Mineral composition

The present study demonstrated that aerobic granular sludge is capable of treating livestock wastewa from a cattle farm in a sequencing batch reactor (SBR) without the presence of support material. A lab scale SBR was operated for 80 d using 4 h cycle time with an organic loading rate (OLR) of 9 kg COD m⁻³ - d⁻¹. Results showed that the aerobic granules were growing from 0.1 to 4.1 mm towards the end of the experimental period. The sludge volume index (SVI) was 42 ml g^{-1} while the biomass concentration in the reactor grew up to 10.3 gL^{-1} represent excellent biomass separation and good settling ability of the granules. During this period, maximum COD, TN and TP removal efficiencies (74%, 73% and 70%, respectively) were observed in the SBR system, confirming high microbial activity in the SBR system. © 2013 Elsevier Ltd. All rights reserved

1 Introduction

In the past, a number of research was performed using aerobic granular sludge in laboratory-scale reactors treating different types of wastewater such as soybean-processing wastewater (Su and Yu, 2005), brewery wastewater (Wang et al., 2007), abattoirs wastewater (Cassidy and Belia, 2005), dairy wastewater (Schwarzenbeck et al., 2005), saline wastewater (Figueroa et al., 2008), and palm oil mill effluent (POME) (Abdullah et al., 2011). Although aerobic granular sludge has been investigated extensively using synthetic wastewater, there is hardly any in aerobic granular sludge under real wastewater.

The growth of Malaysian's livestock industry increased significantly and accelerated gradually each year (Ngo, 2004). In general, livestock wastewater contains high chemical oxygen demand (COD), biological oxygen demand (BOD), color, nitrogen, phosphorus and suspended solids (Lee and Shoda, 2008). High levels of phosphorus and nitrogen in discharging livestock wastewater contribute to eutrophication of receiving wastewater particularly lakes

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and slow moving rivers. Therefore, a specific treatment process is required in treating wastewater

In the past, small quantity of livestock wastewater was discharged without proper technology and contributed substantially to environmental pollution. Most of the livestock industries are employing conventional process for livestock wastewater treatment. The negative impact will be greater if the volume and frequency of the untreated discharged of livestock are higher. Since high costs of the reagent and poor performance of physico-chemical process in removal of soluble organics, biological treatment process are most preferred. Aerobic granular sludge is one of the biological treatment technologies that can successfully cultivated in treating livestock wastewater in a sequencing batch reactor

Therefore, this study attempts to demonstrate the cultivation of aerobic granules applied for the effective bioprocess that is able to treat livestock wastewater. The study focuses on the aerobic gran ular sludge removal efficiency and the effluent quality treating real wastewater. Moreover, biomass profile and settling properties for the treatment of livestock wastewater were also presented in this study. Considering the advantages of aerobic granular sludge over conventional sludge flocs, aerobic granular sludge is a reasonable

Othman et al., 2013



Industrial: Granjas (N y P)

Eliminación DQO: 74% Eliminación NT: 73%

Eliminación PT: 70%

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HIGHLIGHTS

- ▶ Matured aerobic granules were observed in the reactor with the size up to 4.1 mm.
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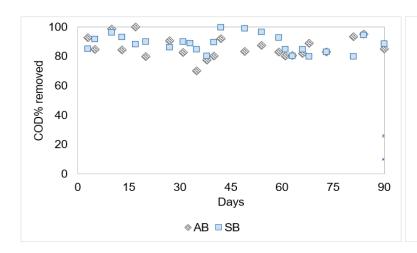
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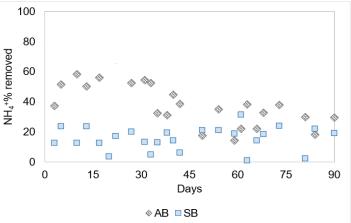
Othman et al., 2013

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Industrial: Granjas





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Amino Acids https://doi.org/10.1007/s00726-022-03168-v

ORIGINAL ARTICLE



Effects of sulphur amino acids on the size and structure of microbial communities of aerobic granular sludge bioreactors

Aurora Rosa-Masegosa ^{1,2} · Lizandra Perez-Bou^{2,3} · Barbara Muñoz-Palazon ^{1,2} · Ontonio Monteoliva-García ⁴ · Alejandro Gonzalez-Martinez ^{1,2} · Jesus Gonzalez-Lopez ^{1,2} · David Correa-Galeote ^{1,2}

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Abstract

Granular activated sludge has been described as a promising tool in treating wastewater. However, the effect of high concentations of sulphur amino acids, cysteine and methionine, in the evolution, development and stability of AGS-SBRs (acrobic granular sludge in sequential batch reactors) and their microbial communities is not well-established. Therefore, this study aimed to evaluate microbial communities' size, structure and dynamics in two AGS-SBRs & do with two different concentrations of amino acids (50 and 100 mg L⁻¹ of both amino acids). In addition, the impact of the higher level of amino acids was also determined under an acclimatization or shock strategy. While N removal efficiency decreased with amino acids the removal of the organic matter was generally satisfactory. Moreover, the abrupt presence of both amino acids developed the removal of the organic matter was generally satisfactory. Moreover, the abrupt presence of both amino acids developed the removal of the organic matter was generally satisfactory. Moreover, the abrupt presence of both amino acids decelerate the removal performance of N, whereas under progressive adaptation, the removal yield was higher. Besides, excelent removal artes of cysteine and methioniane elimination were found, and stages below 80% of the influent studies. Generally considered, the addition of amino acids weakly impacts the microbial communities total abundances. On the contravy, the presence of amino acids sharply modulated the dominant bacterial structures. Furthermore, the highest amino acid concentration under the shock strategy resulted in a severe change in the structure of the microbial community. Acidovarus, Flavobacterium, Methylophilias, Senatorphomonous and Thausers acts odo out as the prominent bacteria to cope with the high presence of cysteine and methionine. Hence, the AGS-SBR technology is valuable for treating influents enriched in sulphur Aa inclusively when a shock strategy was used.

Keywords Cysteine · Methionine · qPCR · Illumina sequencing · N-removal · AGS-SBR

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Aurora Rosa-Masegosa and Lizandra Perez-Bou contributed equally to this article.

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Introduction

The global population is currently increasing exponentially, and it is expected to rise around 30% in the following years, reaching 9.9 billion people by 2050 (Bahar et al. 2020). In addition, many of the essential human activities, such as crop and livestock production, manufacturing of goods, power generation, or domestic activities, rely upon the availability of water in both adequate quantities and of acceptable quality (Huang et al. 2018). Therefore, the increase in the global population is coupled in parallel with higher water requirements (Jones et al. 2021). Concomitantly, sustainable water management is imperative to ensure water availability and easy access worldwide (Mole and Brooks, 2019). On the other hand, the increase in water requirements resulted in the generation of higher quantities of wastewater (WW), making its management a major concern for people's health. Also, it is an environmental hazard due to the ineffective and

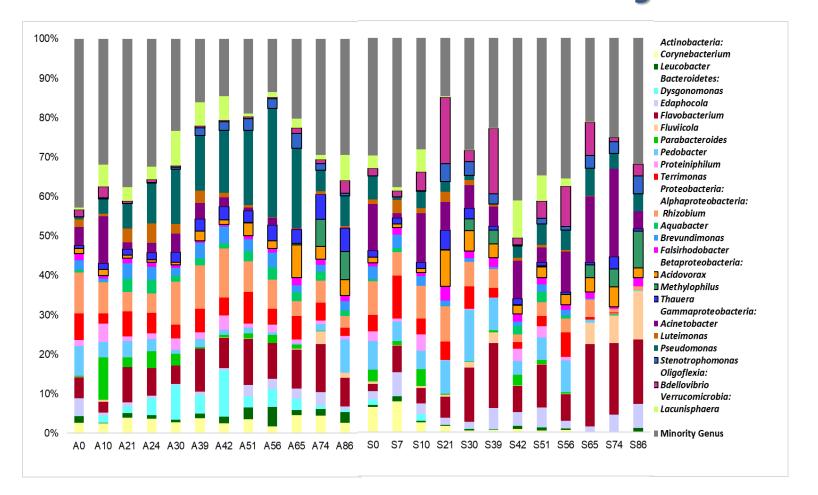
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Rosa-Masegosa et al., 2022



Industrial: Granjas



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ORIGINAL ARTICLE



Effects of sulphur amino acids on the size and structure of microbial communities of aerobic granular sludge bioreactors

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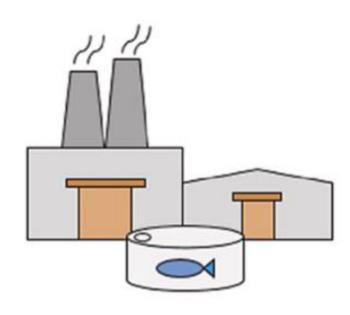
Industrial: Conserveras pescado





Contaminantes

- Materia orgánica
- N
- Alta salinidad
- Grasas





Industrial: Conserveras pescado

Table 1 Average values of medium-low strength wastewater (MS-WW) and high strength wastewater (HS-WW) composition.

<u>Neactor</u> . 3BR, 3000 L.
<u>TRH</u> : 6 h y 15 h (MS-WW y
HS-WW, respectivamente).
<u>Inóculo</u> : fango activo
adaptado a alta salinidad

Panctar CDP 2000 I

Parameter	Unit	MS-WW	HS-WW
Conductivity	mS/cm	22.1 ± 12.5	18 ± 3.6
pH	_	6.26 ± 0.85	6.59 ± 0.23
TSS	g/L	$\textbf{0.250} \pm \textbf{0.137}$	2.37 ± 1.79
VSS	g/L	$\textbf{0.213} \pm \textbf{0.140}$	1.07 ± 0.26
Total COD	mg/L	1203.8 ± 671.9	$10,843.0 \pm 4898.2$
Soluble COD	mg/L	507.4 ± 282.0	7352.2 ± 2227.0
Greases	mg/L	145.7 ± 120.3	1608.2 ± 1208.8
VFAs	mg COD/L	169.7 ± 206.2	3373.8 ± 2295.3
- Acetic acid	mg COD/L	$\textbf{41.4} \pm \textbf{53.1}$	1272.5 ± 814.8
 Propionic acid 	mg COD/L	64.8 ± 11.3	548.2 ± 271.6
 Butyric acid 	mg COD/L	9.2 ± 18.4	1092.5 ± 802.9
- Valeric acid	mg COD/L	20.1 ± 40.0	460.5 ± 417.9
NH ₄ +N	mg N/L	$\textbf{8.18} \pm \textbf{7.45}$	557.7 ± 395.6
TN	mg N/L	21.23 ± 8.36	1025.0 ± 117.4
PO_4^{-3} -P	mg P/L	3.05 ± 2.46	129.0 ± 71.2
Proteins	mg/L	86.46 ± 48.08	2059.6 ± 296.1
Carbohydrates	mg/L	46.28 ± 29.76	514.8 ± 723.7
Chloride	g Cl ⁻ /L	7.37 ± 5.6	$\textbf{4.21} \pm \textbf{1.01}$
Sulphate	$g SO_4^{-2}/L$	1.00 ± 0.77	$\textbf{0.25} \pm \textbf{0.04}$
Sodium	g Na ⁺ /L	4.24 ± 2.90	2.77 ± 0.60
Salinity	g NaCl/L	14.35 ± 5.76	7.41 ± 2.06

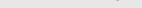
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Journal of Environmental Management





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Features of aerobic granular sludge formation treating fluctuating industrial saline wastewater at pilot scale

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ARTICLE INFO

Industrial wastewater

A pilot-scale sequencing batch reactor, with a working volume of 3 m³, was installed in a fish cannery to develop aerobic granular sludge treating the produced effluents. Depending on the nitrogen (N) and organic matter (COD) concentration, the effluents were named in this study as medium-low-strength (Stage I) and high-strength (Stage II) wastewater. The composition of the wastewater was found to be a crucial factor to select granule forming organisms. With medium-low-strength wastewater as feeding, the first granules were observed after 30 days, but the extremely high COD/N ratios of the wastewater provoked the overgrowth of filamentous bacteria after 4 months of operation (Stage I). When treating high-strength wastewater, stable aggregates with good settleability appeared, but well-shaped granules were not observed since the granulation process was not completed. The system was able to remove both COD (70-95%) and N (30-90%) treating both types of effluents. Biomass growth was the main N removal pathway. The reactor was found to be robust against factory production stops and, thus, a suitable alternative to treat wastewater from industries with discontinuous operation.

Pish canneries can process a wide range of products, and their processing schemes in terms of raw material, source of utility water and unit processes vary among factories (Chowdhury et al., 2010). Fish processing operations usually generate large volumes of wastewater, with a variable composition depending on the raw material and specific processing activity (AMEC Earth and Environmental Limited, 2003). In general, they present high concentrations of organic pollutants in soluble, colloidal and particulate form (1140-90,000 mg COD/L, 20-2000 mg TN/L) and eventually high salt concentration (2-36 g NaCl/L) dhury et al., 2010; Cristóvaō et al., 2015). To protect the environment, this kind of wastewater must be treated previous to its

Biological treatments have been found to be suitable for organic matter and nutrients removal from fish-processing wastewater. Both anaerobic and conventional aerobic systems have been applied for this (mainly anaerobic digesters) present the possibility of energy recovery

by the production of biogas. In addition, they have low operational costs, space requirements and sludge generation. However, parameters, such as ammonia concentration (corresponding to free ammonia co centrations above 500 mg NH3-N/L) or salinity (above 3.5 g Na⁺/L), have been reported to cause the inhibition of the process (Appels et al. 2008). Besides, nitrogen is not removed during anaerobic processes, being necessary an additional unit afterwards to comply with the required effluent quality. Applied aerobic processes, mainly conventional activated sludge reactors and aerated lagoons, can remove nutrients besides organic matter (Chowdhury et al., 2010). However, large spaces for their implantation are required, resulting in high footprints Aerobic Granular Sludge (AGS) is proposed as a promising technol-

ogy to treat this kind of saline industrial wastewater, due to its several advantages compared to other biological processes applied for this purpose. For example, it is a compact technology with high treatmen capacity, compared to activated sludge systems. Besides, the layered structure of the granules enables the simultaneous removal of organic matter and nutrients in a single unit (Nanchar Additionally, it can withstand salt concentrations up to 40 g NaCl/L

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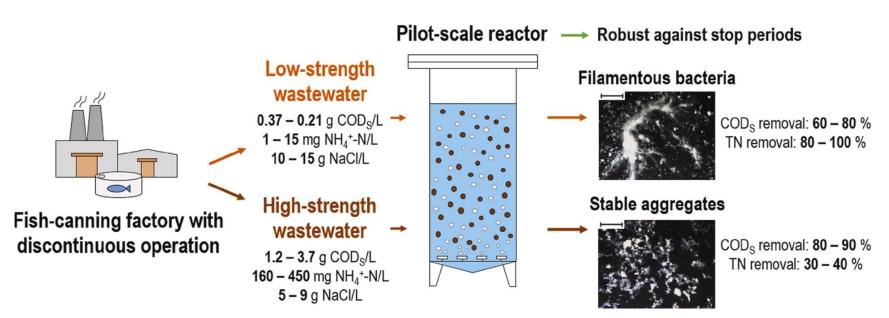
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^{*} Corresponding author.



Conserveras pescado

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Features of aerobic granular sludge formation treating fluctuating industrial saline wastewater at pilot scale



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A pilot-scale sequencing batch reactor, with a working volume of 3 m³, was installed in a fish cannery to develop aerobic granular sludge treating the produced effluents. Depending on the nitrogen (N) and organic matter entration, the effluents were named in this study as medium-low-strength (Stage I) and high-strength (Stage II) wastewater. The composition of the wastewater was found to be a crucial factor to select granule forming organisms. With medium-low-strength wastewater as feeding, the first granules were observed after 30 days, but the extremely high COD/N ratios of the wastewater provoked the overgrowth of filamentous bacteria after 4 months of operation (Stage I). When treating high-strength wastewater, stable aggregates with good settleability appeared, but well-shaped granules were not observed since the granulation process was not completed. The system was able to remove both COD (70-95%) and N (30-90%) treating both types of effluents Biomass growth was the main N removal pathway. The reactor was found to be robust against factory production stops and, thus, a suitable alternative to treat wastewater from industries with discontinuous operation

Pish canneries can process a wide range of products, and their processing schemes in terms of raw material, source of utility water and unit processes vary among factories (Chowdhury et al., 2010). Fish processing operations usually generate large volumes of wastewater, with variable composition depending on the raw material and specific progeneral, they present high concentrations of organic pollutants in soluble, colloidal and particulate form (1140-90,000 mg COD/L, 20-2000 mg TN/L) and eventually high salt concentration (2-36 g NaCl/L) tury et al., 2010; Cristóvaö et al., 2015). To protect the envi-

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Table 3
Fulfilment of the discharge requirements in Stages I and II.

Parameter	Limit concentration	Stage I	Stage II
pН	5.5–9.5	6.5-7.2	7.9-9.1
TSS (mg/L)	125	70-100	200-400
Total COD (mg/L)	350	400-500	400-900
BOD ₅ (mg/L)	150	100-200	40-200
TP (mg/L)	12.5	0.3 - 2.5	50-100
NH_4^+ -N (mg/L)	25	1-6	80-300
TN (mg/L)	57.5	2-8	80-320
Greases (mg/L)	25	70–200	2-10

^{*}The concentrations correspond to the effluent of the pilot plant (outlet of the settler).

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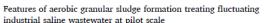


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ABSTRACT

A pilot scale sequencing batch reactor, with a working volume of 3 m², was installed in a fish cannery to develop acrobic granular tudge treating the produced effluence. Depending on the mirrogen (O) and organic matter (CODI) concentration, the effluents were named in this study as medium-low-strength (Stage II) and high-strength (Stage II) and high-strength control of the wasterwater was found to be a crucial factor to select granula-forming organisms. With medium-low-strength wasterwater as feeding, the first granular were observed after 3 days, but the extremely high GODN ratios of the wasterwater provides the evergrowth of filamentous bacteria after of months of operation (Stage II). When treating high-strength wasterwater, stable aggregates with bacteria after of months of operation (Stage II). When treating high-strength wasterwater, stable aggregates with a dependent of the stable of the stable and the stable of the stable and t

Introduction

Fish cannesies can process a voide range of products, and their poscessing schemes in terms of zero material, source of utility water and unit processes vary among factories (Chowollumy et al., 2010). Fish processing operations untally generates large volumes of waterwaters, with a variable composition depending on the zero material and specific processing activity (AMDC Barth and Environmental Limited, 2002), general, they present high concentrations of organic pollutants in soluble, colloidal and particulates from (140-50,000 omg COD/L, 20-200) mg TN/L) and eventually high salt concentration (2-56 g NaC/L) (Chowolluny et al., 2010; Criticoval et al., 2015). De protect the entry-ronment, this kind of wastewater must be treated previous to its discharge.

Biological treatments have been found to be suitable for organic matter and nutrients removal from fish-processing wastewater. Both anaerobic and conventional serobic systems have been applied for this purpose (Chowdhary et al., 2010). Implemented anaerobic processes (mainly anaerobic digesters) present the possibility of energy recovery by the production of biogan. In addition, they have low operational content, space requirements and sludge generation. However, parameters, each, as ammonia concentration (corresponding to free ammonia concentration above 500 mg NHs-N/L) or salimity down 5.8 g Ns⁻²/N, have been reported to cause the inhibition of the process (Appele et al., 2009). Besides, mitosope in not removed during ammentic processes, being necessary an additional unit afterwards to comply with the required effluent quality. Applied aerobic processes, mainly conventional activated sludge reactors and sented lagoons, can remove nutrient besides organic matter (Concombury et al., 2010). However, large spaces for their implantation are required, resulting in high footprints. Aerobic Cranular Sludge (AOS) is proposed as a promaining technical processing and the contraction of the contra

ogy to treat this kind of skilne industrial vastewater, due to its zeveral advantages compared to other biological processes applied for this purpose. For example, it is a compact technology with high treatment capacity, compared to activated kindge systema. Besides, the layered structure of the granules enables the simultaneous removal of organic matter and nutrients in a single unit (Nanchariath and Reddy, 2010). Additionally, it can withstant all eff concentrations up to 9 of 9 NG/LI

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Urbana









Potable

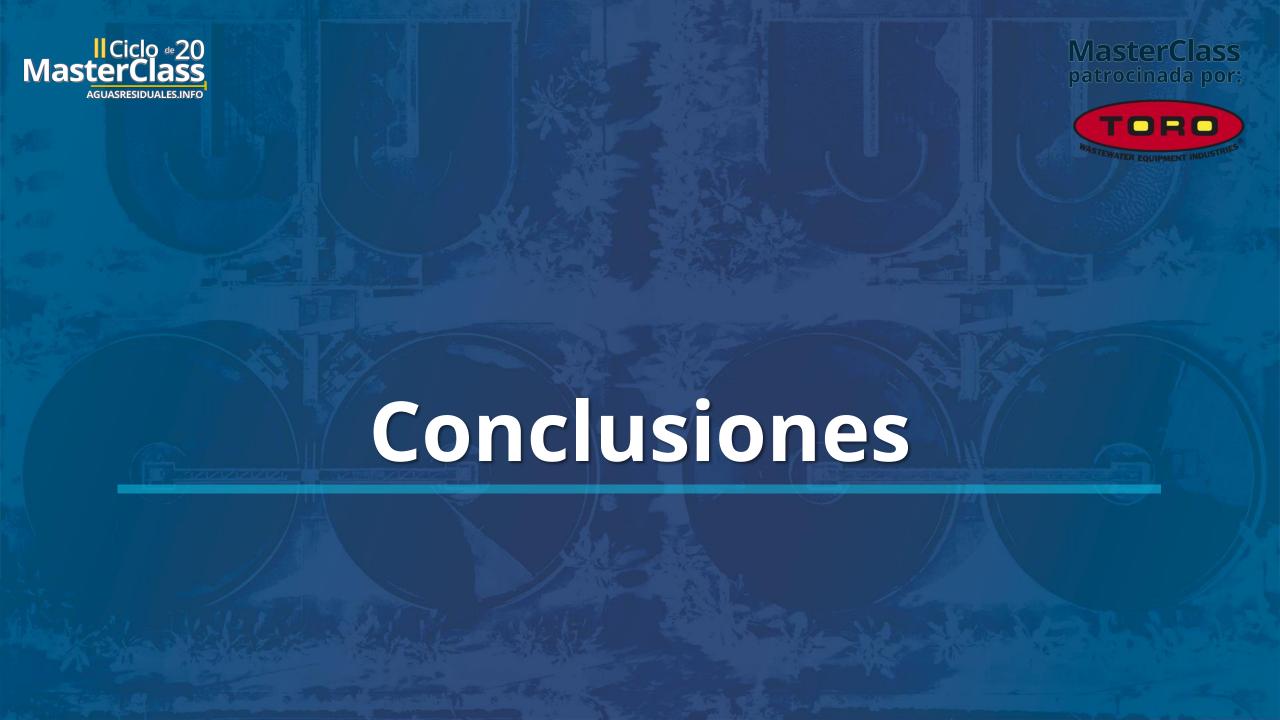






3 SBR Volumen total 4986 L. Torre Cardela (Granada) Eliminación nitrato







Conclusiones

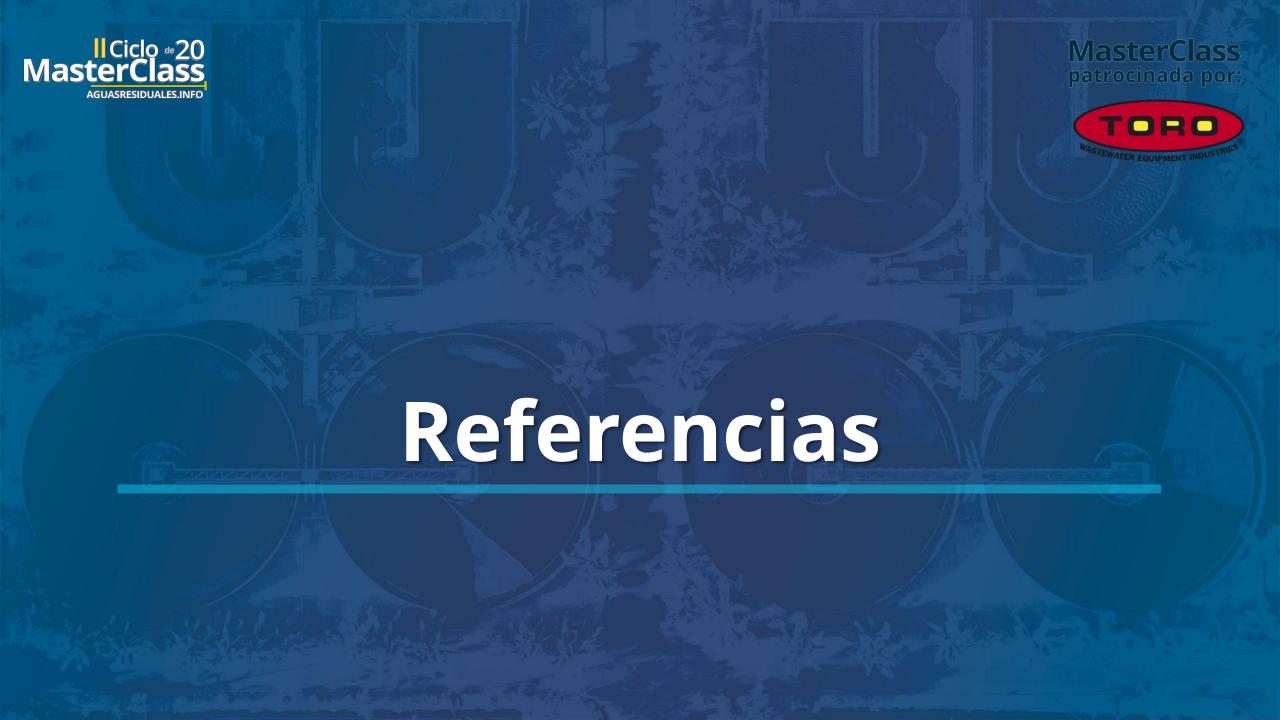




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Colaboración entre Industria-Universidad necesaria.





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Gracias por vuestra atención.

